

USE OF LACCASE-PRODUCING MICROORGANISMS IN MEMBRANE SYSTEMS FOR POLLUTING AGENTS REMOVAL: CONSIDERATION AND PERSPECTIVES

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ABSTRACT

Contamination is defined as the presence of pollution agents in an ecosystem reaching an abnormal concentration level and causing adverse effects. Due to the important environmental problems that these cause in the food chain, diverse methods have been formulated to minimize those effects; among these enzymatic methods associated to microorganisms or other biological agents are used, and systems that allow to metabolize pollutants in an effective and practical way. The combined use of these systems appears to be the most appropriated option to eliminate pollution agents at a reasonable energy cost which allows their use in the industry. The present review identifies the adequate process conditions for technological implementation of biofilm-type membrane systems where laccase-producing microorganisms are used as decontamination agents.

Key words: laccase, biofilm, membrane system, halogen compound.

INTRODUCTION

Contamination is defined as the introduction of some substances or energy forms into an ecosystem that on reaching an abnormal concentration level cause adverse effects. Due to the interaction between the aquatic, earth and atmospheric environments, the circulation of contamination agents has important consequences; the contamination components and their action in the ecosystem which have been comprehensively studied are listed in Table 1.

In the atmosphere, the primary pollution agents are particulate materials (PM), nitrogen oxides (NO_x) sulphur dioxide (SO₂), sulphite reducing gases (TRS), carbon dioxide (CO₂) and carbon monoxide (CO). All of these are involved in the destruction of the ozone

layer when they react with the components of the air. The damage to the protective layer against ultraviolet radiation (UV) is produced by the emission of gases that contain chlorine, forming the chlorofluorocarbons (CFC) which release chlorine atoms (Cl) when they reach the stratosphere, being each of these atoms responsible for the breakup of many ozone molecules (O₃) transforming them into oxygen and chlorine monoxides; chlorine monoxides are removed from the atmosphere during precipitations.

Chlorine reacts with water, forming hydrochloric and hypochloric acids, among others. Ionization of the hydrochloric and hypochloric acids reduces water pH; in the long-term this affects the physiological functions of water organisms. The action of chlorine acids modifies the activity and the quaternary structure of metallothionein (MT). This protein is involved in the transference of metallic ions in water organisms and in the homeostasis of metals at the physiological level; therefore the presence of chlorine acids in estuaries causes metabolic disturbances and increases the respiration rate (Mandel, 2007).

Due to its gaseous nature, Cl is rarely found in its pure form in soil. If it is released into the soil it reacts with water, forming hypochloric acid and hydrochloric acid; these can react with elements in the soil, forming enzyme inhibition compounds which act especially at the extracellular level. Extra-cellular enzymes act as a physical or chemical limit to mineral and organic surfaces. For example, acid phosphatase is one of the many functional phosphatases in the soil and the main agent responsible for the mineralization of organic phosphate compounds in acid soils. An inhibition of the enzyme complex alters the cellular homeostasis of the agents related with the system (Huang, 2003).

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Table 1. Main chemical pollutant agents and their action in the ecosystem.

Chemical compound	Action in the ecosystem	Reference
Chlorofluorocarbons (CFCs)	Break ozone molecules (O ₃) into oxygen and chlorine monoxide.	Darling and Goody, 2007.
Hydrochlorofluorocarbons (HCFCs)	Involved in the transformation of oxygen into carbon monoxide.	McCulloch <i>et al.</i> , 2006; Derwent <i>et al.</i> , 2007.
Nitrous oxide (NO _x)	Generates acid rain, reacts with water in the atmosphere forming sulfuric and nitric acids.	Fang and Mu, 2007.
Polycyclic aromatic hydrocarbons (PAHs)	Formed as a consequence of incomplete combustion of organic matter, they are mutagenic and carcinogenic. The chlorine component enters into the cell membrane structure modifying intra and extracellular regulation functions.	Verdin, 2004; Whiteley and Lee, 2006.
Chlorophenols	Persistent organic compounds, containing one or more chlorine atoms in their molecules. Carcinogenic, they act on the immune, reproductive and nervous systems. At cell level, they cause metabolic disorders and damages in the enzyme system.	Savant <i>et al.</i> , 2006; Choi <i>et al.</i> , 2007; Sahinkaya and Dilek, 2007.

Whether the polluting agents are in the air, water or soil, the problem must be faced in a global manner, since it affects several ecosystems. At the industrial level, the handling of effluents relies on stabilization zones, where the action of microorganisms leads to an oxidation of the pollution component. However, whether the process is anaerobic or aerobic, the system requires a large amount of space and involves important energy and operational costs. Concerning gaseous emissions, the technological alternatives are the destruction of the pollutant, employing chemical and catalytic methods, and the recuperation of contaminated air, which is no more than a series of filtering operations to classify the PM. Besides, there are studies which support the action of certain plants and their associated microorganisms, which exert a mitigating action in the control of atmospheric emissions, to degrade, contain or stabilize environmental contaminants (Montgomery, 2004; Ramos, 2005); this is possible due to the action of the microbiologic-enzymatic systems which involve nitroreductase, dehalogenase, peroxidase, nitrilase and mostly laccase. However, the bioremediation process requires a long time and it is not convenient to couple it to productive processes.

Due to the difficulties mentioned, the idea has arisen to employ enzymes associated to biofilm-type membrane systems, regardless of the polluting agent type of discharge. These systems allow for the transformation of contaminants in an effective manner, at a reasonable energy cost and are suitable for their use in industry. Consequently, the purpose of this article is to review enzymatic methodologies involving laccase-producing microorganisms coupled to an artificial membrane medium as decontamination agent.

Membrane systems associated to laccase-producing microorganisms

Laccase (EC 1.10.3.2) is a polyphenol oxidase which acts on *p*-diphenols; some of these contain halogen elements in their structure (Xu, 1996). Its catalytic site is characterized by four copper (Cu) atoms, which couple to four electrons of the reduction of dioxygen to water in the oxidation of the respective substrate (Solomon and Sundaram, 1996).

There are numerous laccase-producing microorganisms; however, advanced studies in the use of microorganisms attached as laccase producer are focused on a prokaryote belonging to the *Bacillaceae* family, and fungal organisms associated to organic material rotting processes (Table 2), due to their adaptation to variable media and to their capacity of forming a biofilm-type matrix. *Cunninghamella elegans*, *Aspergillus niger*, *Penicillium* spp., *Phanerochaete chrysosporium*, *Pleurotus ostreatus*, *Trametes versicolor* and *Bjerkandera* sp. have shown adaptation to media with a considerable concentration of polycyclic aromatic hydrocarbons (PAHs); they have been able to develop and present a significant potential to metabolize the pollution agent (Cerniglia and Sutherland, 2006).

Membrane systems are structural polymers in charge of providing optimum conditions for the development of the attached microorganisms. Depending on the characteristics of these systems, the action of their enzyme complexes and the removal of the polluting agents in different kinds of effluents are ensured. They are employed as support and development structures, for the microorganism and enzyme production medium

Table 2. Laccase-producing microorganisms.

Microorganism	Type	Reference
<i>Trametes pubescens</i>	White rot fungus	Osma <i>et al.</i> , 2007
<i>Pleurotus ostreatus</i>	Basidiomycete	Prasad <i>et al.</i> , 2005
<i>Cerrera unicolor</i>	White rot fungus	Jarosz-Wilkolazka <i>et al.</i> , 2006
<i>Abortiporus biennis</i>	White rot fungus	Jarosz-Wilkolazka <i>et al.</i> , 2006
<i>Bjerkandera</i> spp.	White rot fungus	Novotny <i>et al.</i> , 2000; Cerniglia and Sutherland, 2006
<i>Coriolopsis</i> spp.	White rot fungus	Novotny <i>et al.</i> , 2000; Cerniglia and Sutherland, 2006
<i>Irpex</i> spp.	White rot fungus	Novotny <i>et al.</i> , 2000; Cerniglia and Sutherland, 2006
<i>Phanerochaete</i> spp.	White rot fungus	Novotny <i>et al.</i> , 2000; Cerniglia and Sutherland, 2006
<i>Pleurotus</i> spp.	White rot fungus	Novotny <i>et al.</i> , 2000; Cerniglia and Sutherland, 2006
<i>Trametes versicolor</i>	White rot fungus	Cerniglia and Sutherland, 2006; Akzu <i>et al.</i> , 2007
<i>Neurospora crassa</i>	Ascomycete	Luke and Burton, 2001
<i>Hirschioporus laricinus</i>	White rot fungus	Banat <i>et al.</i> , 1996
<i>Inonotus hispidus</i>	White rot fungus	Banat <i>et al.</i> , 1996, Kirby, 1999
<i>Phlebia tremellosa</i>	White rot fungus	Banat <i>et al.</i> , 1996, Kirby, 1999
<i>Phanerochaete chrysosporium</i>	White rot fungus	Banat <i>et al.</i> , 1996, Kirby, 1999; Gnanamani <i>et al.</i> , 2006
<i>Funalia trogii</i>	White rot fungus	Ünyayar <i>et al.</i> , 2005
<i>Coriolus versicolor</i>	White rot fungus	Kapdan <i>et al.</i> , 2000
<i>Cunninghamella polymorpha</i>	White rot fungus	Sugimori <i>et al.</i> , 1999
<i>Geotrichum candidum</i>	White rot fungus	Lee <i>et al.</i> , 2000
<i>Rhizopus arrhizus</i>	White rot fungus	Aksu and Tezer, 2000
<i>Fusarium solani</i>	White rot fungus	Verdin, 2004
<i>Dichomitus squalens</i>	White rot fungus	Kunla <i>et al.</i> , 2007
<i>Trichophyton rubrum</i>	White rot fungus	Jung <i>et al.</i> , 2002
<i>Trametes hirsuta</i>	White rot fungus	Rosales <i>et al.</i> , 2007
<i>Chalara</i> (syn. <i>Thielaviopsis paradoxa</i>)	Hyphomycete	Robles <i>et al.</i> , 2002
<i>Bacillus</i> spp.	Bacteria	Dalfard <i>et al.</i> , 2006

and act as selective permeable barriers to the movement of structures or chemical compounds between two phases or conditions. Their geometric configuration depends on the endogenous characteristics of the microorganism (Wang and Wang, 2006); this field is open to far reaching research.

Due to the artificial nature of the membrane, the symbiotic development between the laccase-producing microorganism and the membrane matrix is the subject of this study. On the structural level, the most important characteristics of membrane systems are the morphology, size and density of membrane pores, statistical distribution of pore size, flow turbulence and volume occupied by the pores. These concepts are discussed in Table 3. As regards the functional characterization of the membrane system, what is studied is the permeability of the membrane, retention coefficients, separation

factors, the adsorption characteristics, effective diffusion coefficients and various tests of chemical, mechanical and physical compatibility.

Microbial growth conditions in membrane systems

The growth of enzyme-producing microbial cells in membrane systems can be carried out in aerobic and anaerobic conditions.

The capacity of enzymes to act in metabolic activities involving aerobic conditions is affected by a diversity of factors such as the nutritional composition required by the attached microorganism, the medium pH, the temperature profile, the membrane porosity and particles the retention (Dalfard *et al.*, 2006; Gnanamani *et al.*, 2006; Aksu *et al.*, 2007). The dissolved oxygen level becomes a parameter of the process which determines membrane filterability. At high levels of dissolved

Table 3. The most important structural characteristics of membrane systems associated to laccase-producing microorganisms.

Structural characteristic	Justification	Reference
Morphology and average size of the pores	Generally expressed as a form factor and a radius or diameter value of equivalent pore	Wang and Wang, 2006; Vladislavljević <i>et al.</i> , 2006
Surface density of the pores	Distribution of the number of pores in area unit of a membrane system	Tran <i>et al.</i> , 2006
Volume of porosity	Fraction of the total volume of the membrane which is occupied by pores or holes	Khayet, 2003; Wang and Wang, 2006
Tortuosity	Pores are not generally cylindrical, so the area occupied at the surface does not correspond to the volume occupied within the membrane	Li and Wang, 2006; Sun <i>et al.</i> , 2006
Statistical distribution of pore sizes	There is no uniformity in the physical characteristics of the pores and their distribution within the membrane system	Khayet, 2003; Wang and Wang, 2006

oxygen, the resistance in the membrane increases, this reflects in the porosity of the membrane and its capacity to retain the particle material of the effluent (Jin *et al.*, 2006).

There are few studies involving the development of biological organisms producing enzymes for the removal of polluting compounds in aerobic conditions. In the removal of polluting agents by laccase, the reactive oxidation of the polluting agent is less effective in aerobic conditions due to the characteristics of production of the enzyme.

The mechanisms involved in anaerobic processes are reductive dechlorination, where a halogen group is replaced by a hydrogen group, involving the transfer of two electrons, and dehalorespiration, a mechanism where a chlorinated hydrocarbon is employed as electron acceptor to aid microbial growth.

Both aerobic and anaerobic metabolic pathways, are efficient for the removal of polluting compounds; however, there are direct and indirect advantages that point to a process without an external electron acceptor. When oxygen appears as the sole electron acceptor, dissolved oxygen is a control parameter of the process (Jin *et al.*, 2006); the disposition of the stabilization zones is more demanding in aerobic processes, either due to physical management or to raw material requirements (Savant *et al.*, 2006) and the aerobic processes involve the volatilization of organic compounds, this implies another phase of the process (Kosaric and Blaszczyk, 1991). On the contrary, sub-products of anaerobic conditions, with high caloric content, can be re-circulated to other phases of the productive process (Rajeswari *et al.*, 2000).

Types of reactors associated to a membrane matrix and operating modes

The reactors involving the use of a permeable membrane of the biofilm type are of the packed-bed (Canovas-Diaz and Howell, 1988), which in addition employ natural or synthetic materials as support for the membrane (Table 4). The process begins with a flow of water followed by a counter-current flow of air. The microorganisms attached to the membrane matrix develop until they reach a stabilization phase, where an efficient removal of the noxious agents takes place. Some drawbacks may arise in the operation of this type of reactor, such as a low degradation rate and the heterogeneity of the membrane during the process. This brings about an accumulation of polluting agents in high concentrations in the membrane structure, and this is noxious to the attached microorganisms. The affinity of the enzyme for the substrate is also relevant in the removal of polluting agents by means of packed bed reactors; laccase has a high affinity to phenolic and p-diphenolic compounds (Robles *et al.*, 2002); however, enzyme inhibitors modify enzymatic activity, resulting in an inadequate action against polluting agents; particularly, the presence of compounds containing Fe_2^+ , Ag_2^+ , and Cu_2^+ affects the formation of the enzyme substrate complex, causing a low degradation rate (Xu, 1996; Robles *et al.*, 2002).

A configuration which considerably improves mass transfer is to use membranes as independent barriers; these act in a counter-current of air and water (Figure 1). The geometric configuration of the membrane must be defined, as it must be fitted to the characteristics of the microorganism. The ideal strategy is to use tubular structures, because the contact surface is greater than in

Table 4. Support materials used in reactors associated with membrane matrix.

Natural materials	Reference
Soil materials	Smet <i>et al.</i> , 1996
Compost	Cho <i>et al.</i> , 2000
Moss soil materials	Yang and Allen, 1994
Wood shavings	Cho <i>et al.</i> , 2000
Rock's zones with low levels of silica	Nukunya <i>et al.</i> , 2005
Synthetic materials	
Ceramic support	Gabriel and Deshusses, 2003
Polyethylene filter	Sorial <i>et al.</i> , 1998
Polyurethane filter	Koe <i>et al.</i> , 2001
Foam	Kinney <i>et al.</i> , 1996
Granular activated carbon	Ho <i>et al.</i> , 2007
Extruded diatomaceous soil pellets	Kim <i>et al.</i> , 2005; Kim and Sorial, 2007

a spherical configuration. As the contact surface between the membrane system and the polluting agent increases, the mass transfer indexes improve considerably due to the extracellular secretion of laccase (Luke and Burton, 2001). The configurations for anaerobic processes are similar. The only difference is the absence of air flow (Rajeswari *et al.*, 2000).

In the analysis of mass transference for any configuration of reactor associated to membrane structures, it must be considered that: (1) the transfer of flows in aqueous or liquid phase must be mainly diffusive, since this allows partial control of the development conditions of the attached microorganism. At the intra-particle level, the nature of the substrate is mostly insoluble but microorganisms use the soluble phase of the substrate for their development. However, this mechanism must be evaluated because this process takes time; (2) the resulting metabolites must be removed quickly and efficiently, the use of forced flows appears as the best option; (3) the porosity of the membrane is an ambiguous factor, on one hand it has a negative influence on the metabolic processes inside the membrane structure (Raghavarao *et al.*, 2003); however a degree of membrane porosity facilitates diffusion processes (Miller and Grant, 2005; Jin *et al.*, 2006); (4) wear of the membrane lowers the efficiency of the process.

Likewise, the interaction between flow, membrane system and catalytic reaction has energetic implications which result in a global heat transference in the process, as heat loss or gain. However, the action of the attached

microorganisms is sensitive to this phenomenon so it must be taken into account that: (1) the incoming and outgoing flows from the bioreactor module are the consequence of perceptible energetic changes and of the concentration of products such as O₂, CO₂, and water; (2) natural convection or diffusion mechanisms must be employed for the control of non-forced flows of gaseous phase components; (3) special care must be given to the interaction between the material of the bioreactor and the currents flowing around it. It is recommended to use convective methods for the cooling of those flows.

The configuration of reactors on pilot and laboratory scales are mostly continuous sequences of reactors, counter-current reactors, anaerobic contact filters and membrane-granular activated carbon matrix diphasic reactors (Sahinkaya and Dilek, 2007).

The contaminants removal level does not depend only in the control of the effluent; an adequate characterization of the microorganisms to be employed must be made and most particularly, of their enzyme production rate; this was demonstrated by Choi *et al.* (2007), who found that, in spite of having effected a good control of the effluents, the removal levels of mono, di and trichlorophenols were not adequate due to the toxicity of the compound for the microorganisms and an inadequate enzyme identification. Sahinkaya and Dilek (2007) determined that the acceptable chlorophenol removal levels are of 53 mg L⁻¹ for monophenols and 25 mg L⁻¹ for dichlorophenols.

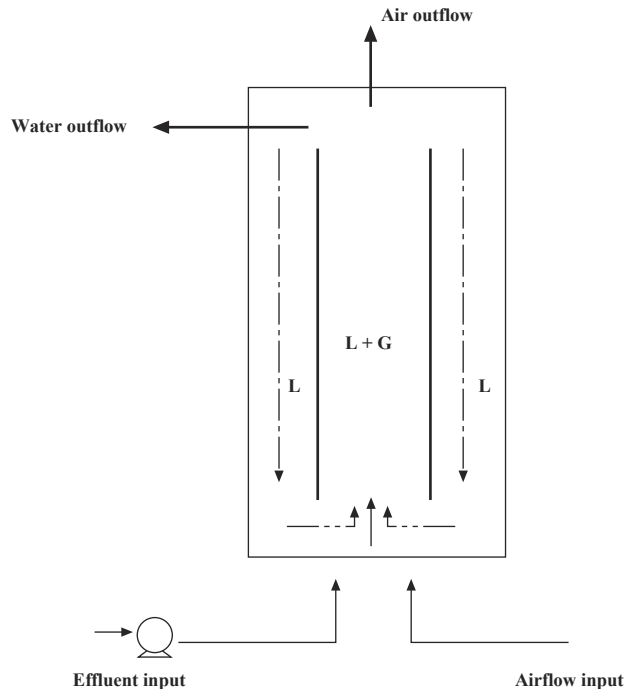


Figure 1. Reactor with immobilized membrane. L: liquid, G: gas.

Source: Rajeswari *et al.*, 2000.

The configuration of the reactor and the support material of the membrane are also key factors of the membrane systems. For example, Alvarez (2006) using pumice stone as membrane support in a fluidized bed reactor operated the system during 300 days. This is double the time than when using conventional methodologies applied in the laboratory, which reach an acceptable operation time of 40 to 120 days.

An alternative in the research of chlorinated effluents handling is the use of solar radiation and photo catalysis, as an adjunct treatment to the membrane systems with laccase-producing microorganisms (Malato, 2003; Pedroza, 2007); after the biological action of laccase, the photo catalytic treatment mineralizes the chlorinated compound. This alternative is a promising study field, due to the increase in the production of photocells, a component of high economic value.

Aerobic and anaerobic membrane systems are also combined for the removal of polluting agents in water (Tartakovsky, 2005). The experimental method evaluates control of trichloroethylene (TCE), making use of the measured synergic action of methanogenic archaea in aerobic and anaerobic phases. Although the parameters of reference mentioned at the laboratory

level are TCE removal levels and mass transference, the characteristics of the attached microorganisms are left in secondary importance.

In the processes of volatile compounds removal (VCR), the control options are centered in three technological solutions involving the use of microorganisms and the production of enzyme complexes. These are the biocleaner, bioregulating filter and biofilter (Garner and Barton, 2002). The process is based on aerobic metabolism, demanding a high humidity content to improve the temperature profile of the process and the transfer of oxygen is fundamental.

The biocleaning system includes airflow recirculation through two modules: a conventional filter to standardize the particle size and a membrane module where the oxidation of the volatile organic compounds takes place (Figure 2). The membrane system has a hydrophobic phase, which regulates the contact of the microorganism with the air flow (Freitas dos Santos *et al.*, 1997).

The biofilter is a long duct, with nutrients for the development of microorganisms and with support materials backing the process in the transference of air flows (Nukunya *et al.*, 2005; Bhat *et al.*, 2006). There are

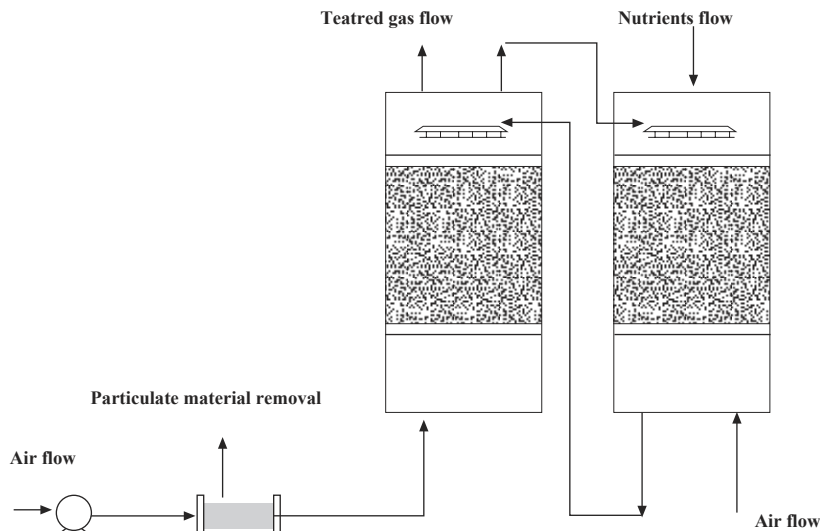


Figure 2. Bioscrubber system with flow recirculation in the modules of conventional filter and reactor.

Source: Adapted from Freitas dos Santos *et al.* (1997).

two phases of preparation of the flow before the contact with the air and the membrane system: a conventional filter associated with activated carbon to reduce the concentration of pollution agents and a humidification module to increase the humidity content and cool the air current. Two operations are identified in the membrane module: (1) the transference of polluting agents from the gaseous to the aqueous phase, which depends mostly on the characteristics of the substrate in the membrane structure, flow condition of the aqueous phase, nature of the transference of agents, possible recirculation of the same in the flows converging in the process and diffusion values of the pollution agents (Jin-Ying *et al.*, 2005); (2) the oxidation of the contamination agents by the attached microorganisms.

The biosifier filter operates on the same parameters as the previous process. However, it differs in the humidification phase of the airflow. The membrane system that comes into contact with the airflow has an adequate humidity content to favor the contact and action of the attached microorganism. Humidification takes place by counter-current trickling (Soccol *et al.*, 2003). The system is similar to the operation of a gas absorption tower (Figure 3). A drawback of the biosifiers is the control of the convective and conductive heat transference through the membrane system, if this is not adequate, the development of the microorganisms would be affected. This aspect is not often reported in scientific literature.

Fields of exploration for the application of membrane systems in the removal of pollution agents

The literature examines several fundamental aspects for the investigation of membrane systems, these are: the use of transgenic lines, strategies to increase enzyme production yields, modeling of enzyme production kinetics and mass transfers, and the assessment of different substrates.

In order to model the kinetic processing of a pollution agent, a wide knowledge of the following aspects is required: (1) diffusion coefficients of the pollution agent in the aqueous, air and atmospheric environments; (2) conditions associated to the environment such as temperature and pH, and (3) concentration of the pollution agent.

On the other hand, filtering always appears as a previous step to the membrane system and the characteristics of the filter to remove pollution agents is another aspect susceptible of study. To use soil materials favorable to the development of microorganisms ensures the catalytic action on the noxious compound and so the toxicity level is drastically reduced, assuring a longer operational time in the bioreactor phase. An alternative is to use soil substrates of mineral type (Gadd, 2007) which allow the development of microorganisms, mostly fungi, which high enzyme production levels.

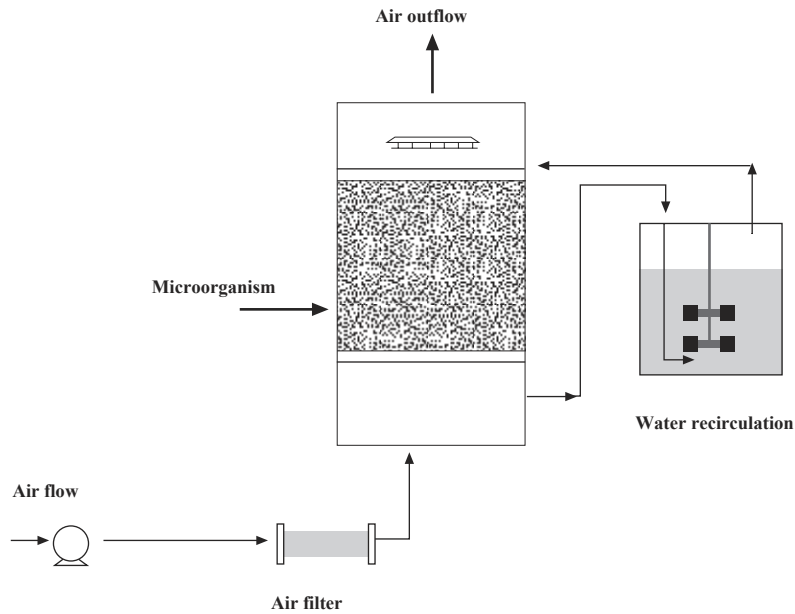


Figure 3. Biodispenser with a recycling of water for humidification processes

Source: Adapted from Soccol *et al.*, 2003.

CONCLUSIONS

To employ membrane systems with laccase-producing microorganisms and obtain their development as they come into contact with liquid or gaseous flows containing compounds of a high degree of toxicity, especially of halogen elements, has turned into a promissory technology with a wide range of application; however, the problems to put the technological tools into operation on an industrial scale are still considerable. Membranes can vary significantly in their structure and constitution and their functional behavior is therefore different; any change can modify their structure and consequently reduce the efficiency of contaminants removal. For this reason the membrane materials, their behavior in the bioreactor and their affinity with the attached microorganism, are fields open to further research.

RESUMEN

Empleo de microorganismos productores de laccasa en sistemas membranarios para remoción de agentes de polución: consideraciones y perspectivas. Liliana Serna C.^{1*}, y Fernando L. Cuesta A.¹ La contaminación se define como la presencia en un ecosistema de agentes de polución que alcanzan un nivel de concentración considerablemente alto, ocasionando efectos adversos. Ante los grandes problemas ambientales que éstos ocasionan en la cadena alimentaria, se han formulado diversos métodos para mitigar los daños, entre los que se encuentran el empleo de métodos enzimáticos asociados a microorganismos o a otros agentes biológicos, y sistemas que permiten el metabolismo de contaminantes de forma efectiva y práctica. Utilizar de manera conjunta estos sistemas, aparece como la opción más acertada para eliminar agentes de polución a un costo energético razonable, lo cual permite emplearlos en la industria. En la presente revisión se identifican las condiciones de proceso propicias para la implementación tecnológica de sistemas membranarios tipo biopelícula que utilizan microorganismos productores de laccasa como agentes de descontaminación.

Palabras clave: laccasa, biofilm, sistema membranario, compuestos halogenados.

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ERRATUM

In journal Vol. 68 Nº 3:

On p. 221, please add in the footnote the authors' affiliation: Dr. Reinaldo Campos-Vargas and Dr. Bruno G. Defilippi: "Plant Cell Biotechnology Millenium Nucleus".

On p. 230, the last line before RESULTS says "Figures 1 to 4"; it should read "Figures 1 to 3".

On p. 284, INTRODUCTION, line 4 says "*Prunas avium*"; it should read "*Prunus avium*".

On p. 295, Palabras clave, must say "residuos de té".